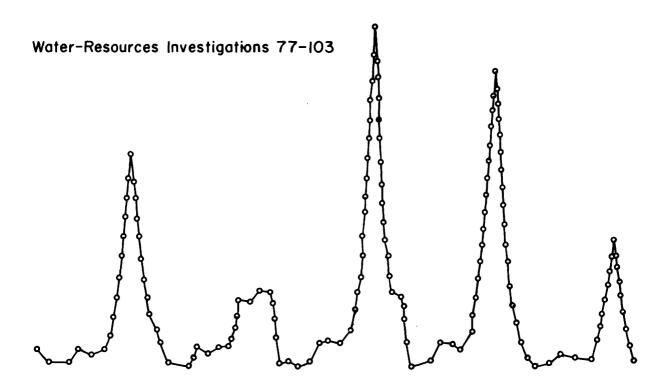
AN ANALYSIS OF SALINITY IN STREAMS OF THE GREEN RIVER BASIN, WYOMING

U.S. GEOLOGICAL SURVEY





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AN ANALYSIS OF SALINITY IN STREAMS OF THE

GREEN RIVER BASIN, WYOMING

By Lewis L. DeLong

U.S. GEOLOGICAL SURVEY

Water-Resources Investigations 77-103



September 1977

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CECIL D. ANDRUS, Secretary

GEOLOGICAL SURVEY

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CONVERSION FACTORS

Multiply English units	<u>By</u>	To obtain metric units
cubic feet per second (ft ³ /s)	0.02832	cubic meters per second (m ³ /s)
tons per day	.9072	megagrams per day (Mg/d)

AN ANALYSIS OF SALINITY IN STREAMS OF THE GREEN RIVER BASIN, WYOMING

by Lewis L. DeLong

ABSTRACT

Dissolved-solids concentrations and loads can be estimated from streamflow records using a regression model derived from chemical analyses of monthly samples. The model takes seasonal effects into account by the inclusion of simple-harmonic time functions. Monthly mean dissolved-solids loads simulated for a 6-year period at U.S. Geological Survey water-quality stations in the Green River Basin of Wyoming agree closely with corresponding loads estimated from daily specific-conductance records. In a demonstration of uses of the model, an average gain of 114,000 tons of dissolved solids per year was estimated for a 6-year period in a 70-mile reach of the Green River from Fontenelle Reservoir to the town of Green River, including the lower 30-mile reach of the Big Sandy River.

INTRODUCTION

Water demands in the Green River Basin of Wyoming (fig. 1) are increasing as a result of existing and potential development of extensive coal, oil, gas, uranium, and oil-shale resources. Planners need more useable information than is now available concerning the effects of proposed development alternatives on the water resources of the basin (Lowham and others, 1976). Water quality, specifically salinity, is an important factor in determining water use and in assessing possible impacts of those uses over time. Salinity data have been collected on the Green River and its major tributaries during the last 25 years, but use of the data in tabular form as published has been limited. A quantitative description of salinity in the Green River and its major tributaries is useful to the evaluation of alternative development plans.

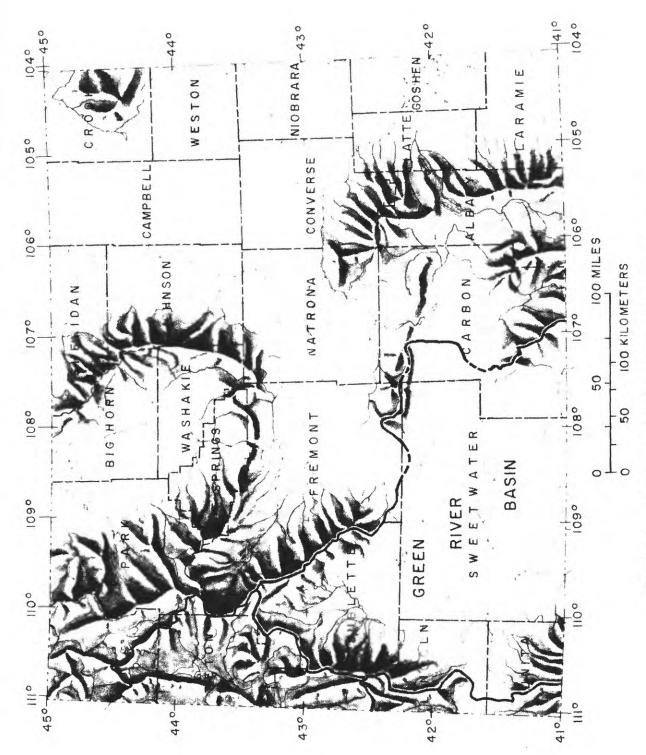


Figure I.—Location of Green River Basin in Wyoming.

Purpose

The purpose of this report is to present a method for converting salinity data into more useable information. Specific objectives are to develop and demonstrate a regression model that would enable daily concentrations and monthly and annual mean loads of the major dissolved inorganic constituents to be estimated at streamflow stations where only monthly samples have been collected.

Data Analyzed

Data analyzed in this report are from streamflow and water-quality stations operated by the U.S. Geological Survey in cooperation with other Federal agencies and with the State of Wyoming. Station locations are shown in figure 2. Table 1 lists sampling stations and period of record for which data were analyzed. In general, the data include analyses of the major inorganic constituents from discrete samples collected before October 1975. Several of the stations have historical water-quality records available in addition to the data used for this study.

METHOD OF ANALYSIS

A quantitative description of the solutes transported in a stream system is useful to evaluate the impacts of proposed or past surface—water development projects (such as reservoirs, irrigation systems, and withdrawals for municipal or industrial use). Many published water—quality records consist of analyses of monthly samples. Natural varia—bility of streamflow during a month reduces the value of a discrete sample to represent streamflow quality throughout the entire month. When daily streamflow records are available in addition to monthly water—quality records, an improved representation of streamflow quality throughout the month may be obtained from estimates that utilize functional relations between streamflow and solute concentration. Multiple—variable regression is used in this report to define the relation of solute concentration to streamflow and day of the year.

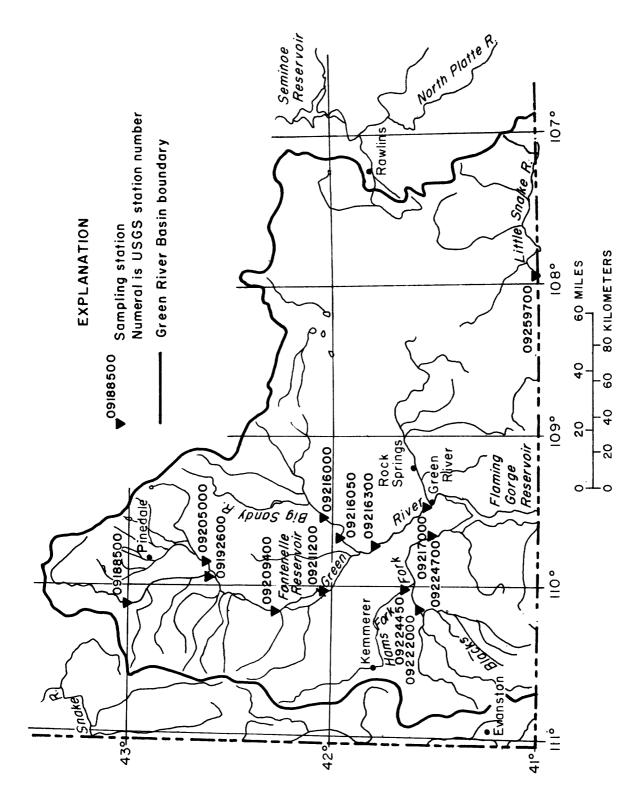


Figure 2.—Locations of quality-of-water sampling stations.

Table 1.--Chemical quality-of-water sampling stations in Wyoming

[See fig. 2 for locations.]

Station	Name	Water Years analyzed
09188500	Green River at Warren Bridge, near Daniel	- 1968–75
09192600	Green River near Big Piney	- 1967–75
09205000	New Fork River near Big Piney	- 1969–75
09209400	Green River near La Barge	- 1970–75
09211200	Green River below Fontenelle Reservoir	- 1970–75
09216000	Big Sandy River below Eden	- 1961–75
09216300	Green River at Big Island, near Green River-	- 1966–75
09217000	Green River near Green River	- 1969-75
09222000	Blacks Fork near Lyman	- 1970-75
09224450	Hams Fork near Granger	- 1969–75
09224700	Blacks Fork near Little America	- 1970–75
09259700	Little Snake River near Baggs	- 1965–74

Wyoming Department of Environmental Quality.

Wyoming Department of Agriculture.

U.S. Geological Survey.

U.S. Bureau of Reclamation.

Cooperators: U.S. Bureau of Land Management.

Two-Variable Regression Model

Dissolved-solids concentration in a stream is related to many factors, but one of the most important is the volume of water available for dilution (Hem, 1970, p. 271). In general, higher concentrations occur at lower streamflows, and with increasing flows concentrations tend to decrease.

Concentration of the major dissolved inorganic constituents in a stream can be related to streamflow by the following two-variable regression equation (Steele, 1973 and 1976):

$$C = A O^{B}$$
 (1)

where

C = concentration, in milligrams per liter,

Q = streamflow, in cubic feet per second, and

A and B = regression coefficients.

An example of this relation is shown in figure 3. Concentration residuals (differences between estimated and observed concentrations) shown in figure 4 are consistently positive during some periods and negative during other periods.

Seasonal shifts in the concentration-flow relation, as exemplified in figure 4, are typical of data for the stations analyzed in this report and, in general, limit the application of equation 1 to depict concentration-flow relations in the Green River Basin. The regression procedure, assuming a constant year-round relation, causes concentration residuals totaled over the entire regression period to approach zero. Positive residuals during one period are balanced against negative residuals during another period. Because streamflow is not evenly distributed over time, residuals from loads calculated from streamflow records and estimated concentrations will normally not approach zero. This leads to inaccuracies both in estimation of annual loads at a given site and seasonal distribution of the annual load over the year.

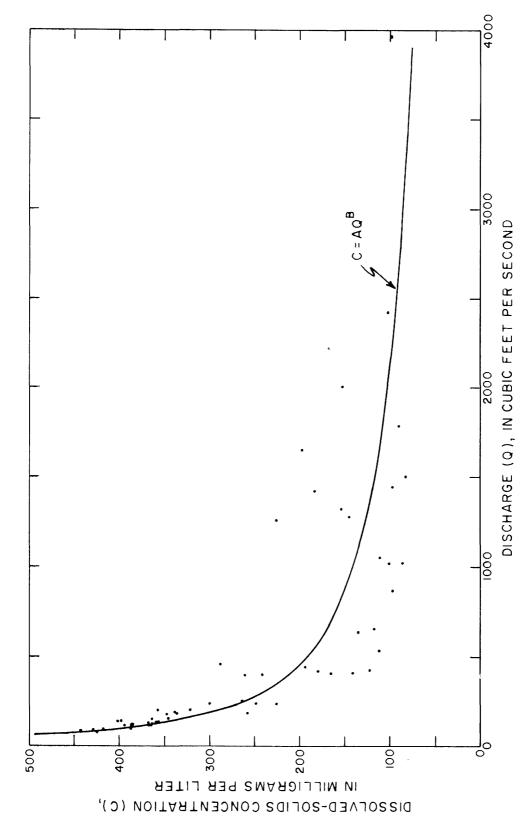
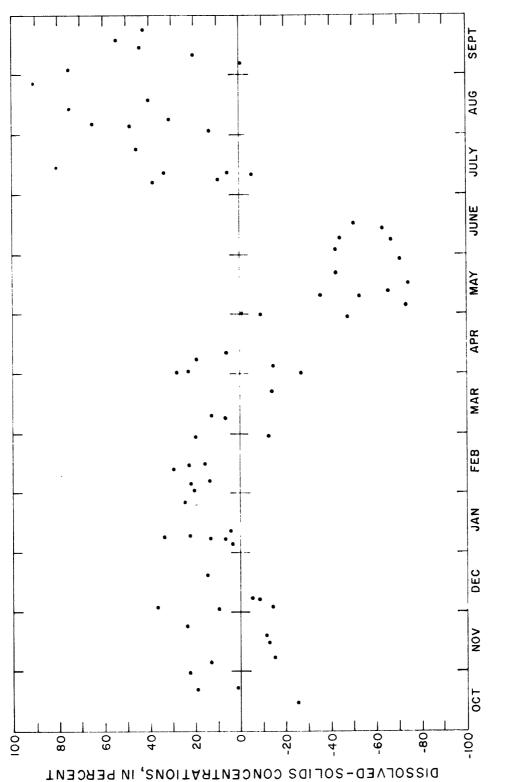


Figure 3 —Relation of dissolved solids to discharge at station 09188500 Green River at Warren Bridge, near Daniel, Wyoming.



DIFFERENCE BETWEEN ESTIMATED AND OBSERVED

Figure 4.—Difference between dissolved-solids concentrations estimated from the two_varible-regression model and dissolved-solids concentrations observed at station 09188500 Green River at Warren Bridge, near Daniel, Wyoming.

Multiple-Variable Regression Model

Seasonal effects were accounted for quantitatively by adding a season-related variable to the regression model. Water temperature would have been an obvious selection for this variable, but time, expressed as day of the water year (table 2), was used because it simplified later simulation efforts. Seasonal effects are incorporated into coefficients A and B in equation 1 by using the following functions:

$$Log_{10} A = B_0 + B_1 sin(\alpha t) + B_2 cos(\alpha t) and$$
 (2)

$$B = B_3 + B_4 \sin(\alpha t) + B_5 \cos(\alpha t), \qquad (3)$$

where

t = day of the water year (table 2),

 α = 0.987 degrees per day or 0.0172 radians per day, and

 B_0 through B_5 = regression coefficients (table 3).

Parameters B_0 through B_5 were determined for the major dissolved inorganic constituents by a multiple-variable regression technique using a computer program developed by K. C. Glover (written commun., 1976). Regression-analysis results for the stations covered in this study are listed in table 3. To demonstrate the improved accuracy of the multiple-regression model in describing variability of dissolved-solids concentration, residuals of the model are plotted versus time in figure 5. This may be compared with the previous two-variable regression example (fig. 4). The same data for station 09188500 were used in both cases. Similar changes in terms of reduced magnitude and more random time-series distribution of residuals were found for other stations analyzed in this report. Hence, subsequent computations in this report utilize the model determined by the multiple-regression technique.

DIFFERENCE BETWEEN ESTIMATED AND OBSERVED

multiple-variable-regression model and dissolved-solids concentrations Figure 5.—Difference between dissolved-solids concentrations estimated from the observed at station 09188500 Green River at Warren Bridge, near Daniel, Wyoming.

APPLICATION OF THE MULTIPLE-VARIABLE REGRESSION MODEL

Computation of Monthly Mean Dissolved-Solids Loads

Monthly mean dissolved-solids loads can be computed from daily streamflow records using the multiple-variable regression model previously described to estimate daily concentrations in the following relation:

$$\overline{L} = (b/d) \sum_{j=1}^{d} C_{j} Q_{j}$$
(4)

where \overline{L} = monthly mean load, in tons per day,

b = 0.0027 (tons per day (milligram) (cubic feet per second),

d = days per month,

j = day of month,

C_j = daily concentration, in milligrams per liter, and

Q; = daily discharge, in cubic feet per second.

Daily dissolved-solids concentrations can be estimated by another method when daily specific conductance data are available. Dissolved-solids concentration can be related to specific conductance (fig. 6) by the following equation (Steele, 1973):

$$C_{\dagger} = E + F K \tag{5}$$

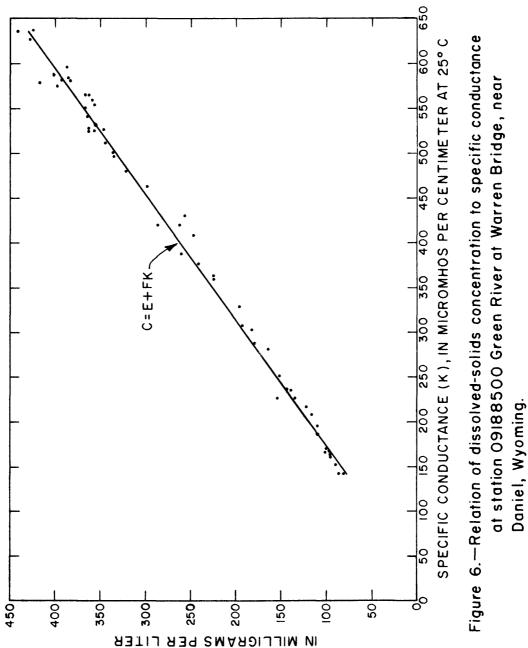
where C_j = daily dissolved-solids concentration, in milligrams per liter,

E and F = regression coefficients (table 4), and

K = specific conductance, in micromhos per centimeter at 25°C.

Because of the large number of calculations involved in estimating dissolved-solids loads, equations 1 through 5 were incorporated into a computer program developed by K. C. Glover (written commun., 1976). Card output from the program was used with an off-line card reader and X-Y plotter to produce solute-load hydrographs as exemplified for dissolved solids in figures 7-14. Individual constituent concentrations and loads (table 3) also can be estimated and plotted.

Semi-quantitative conclusions can be drawn from the dissolved-solids-load hydrographs (figs. 7-14). At stations 09209400, 09211200, 09217000, 09222000, and 09224700 (figs. 9, 10, 12, 13, and 14) where daily specific-conductance data are available, generally good agreement exists between loads computed from concentrations estimated by the two previously described methods. Because identical scales are used on the



DISSOLVED-SOLIDS CONCENTRATION (C),

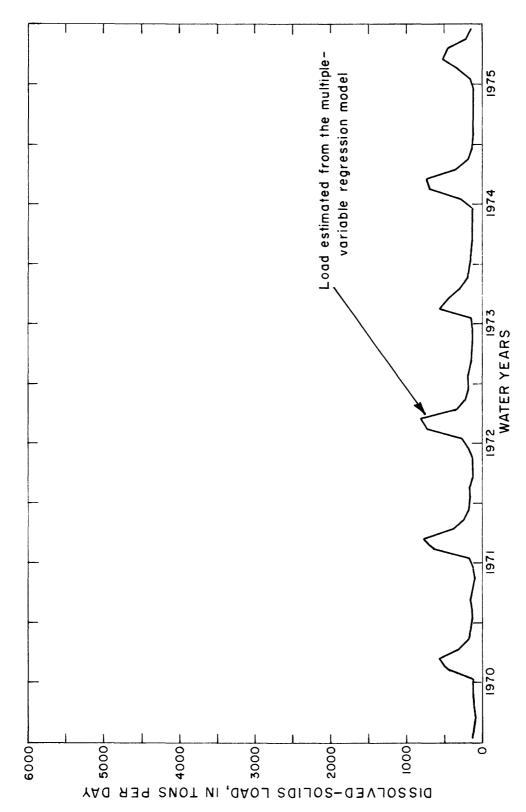


Figure 7.—Monthly mean dissolved-solids loads at station 09188500 Green River at Warren Bridge, near Daniel, Wyoming.

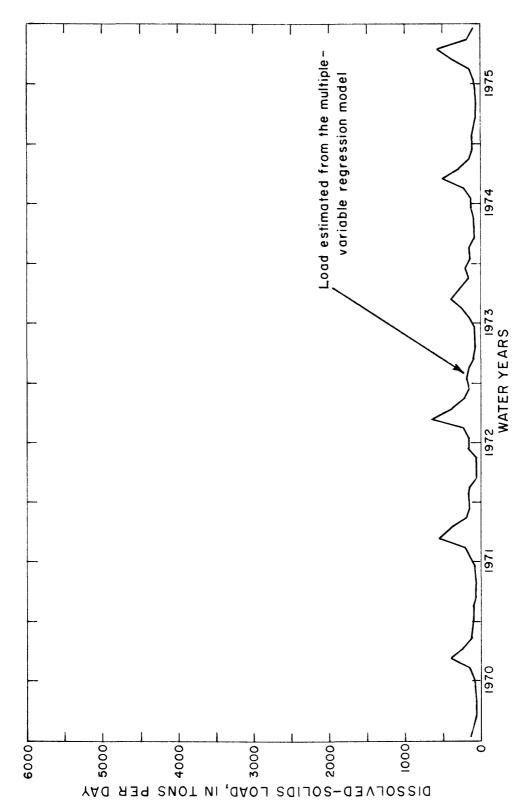


Figure 8.—Monthly mean dissolved-solids loads at station 09205000 New Fork River near Big Piney, Wyoming.

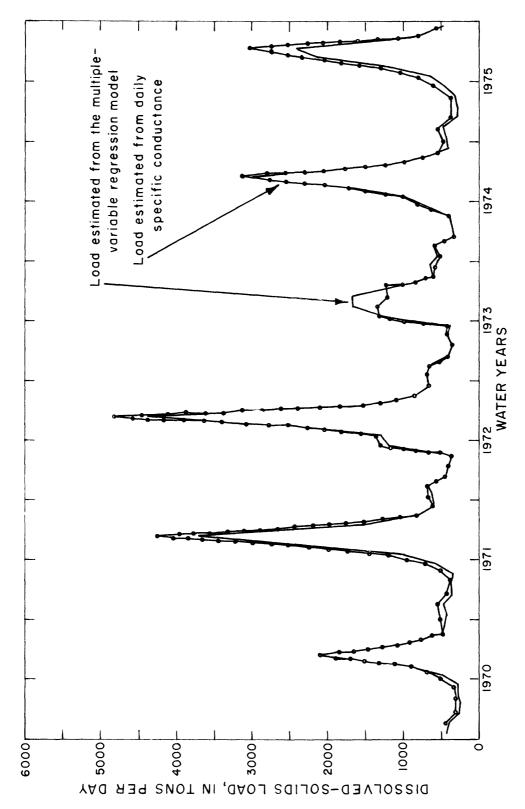


Figure 9.—Monthly mean dissolved-solids loads at station 09209400 Green River near La Barge, Wyoming.

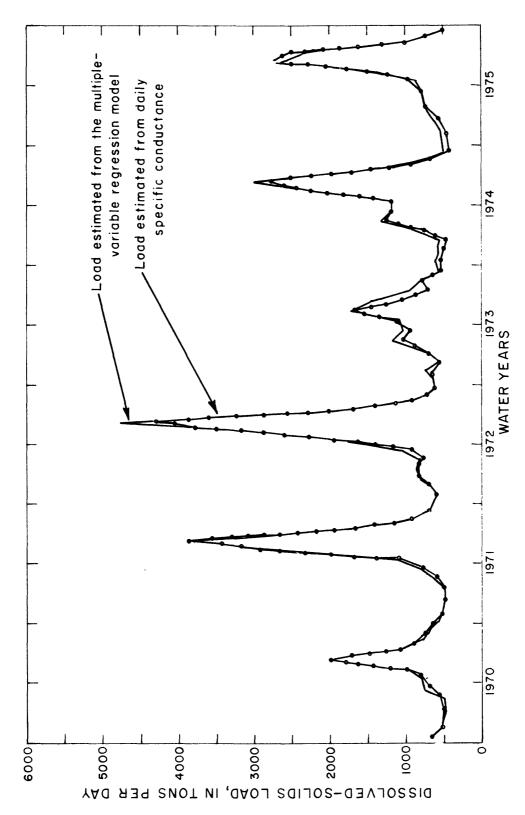


Figure 10.—Monthly mean dissolved-solids loads at station 09211200 Green River below Fontenelle Reservoir, Wyoming.

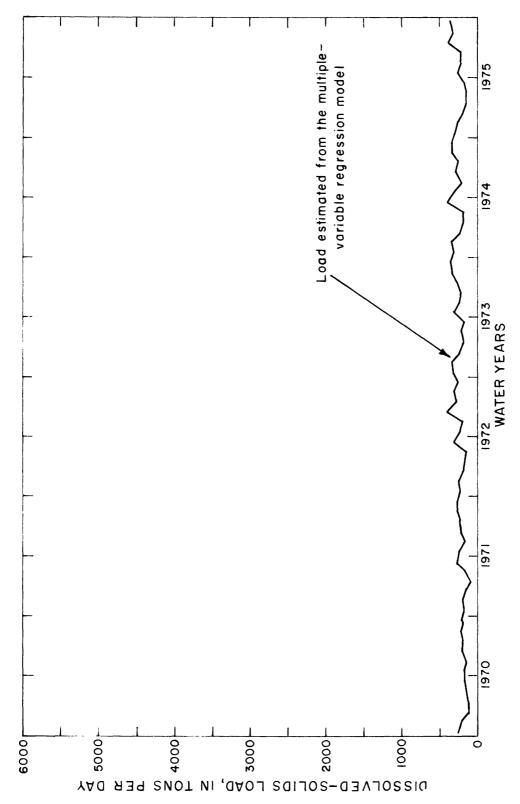


Figure 11.—Monthly mean dissolved-solids loads at station 09216000 Big Sandy River below Eden, Wyoming.

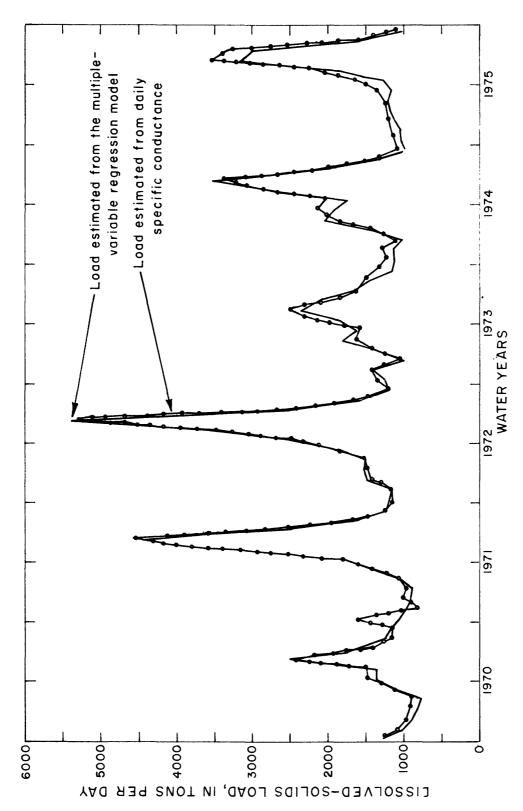


Figure 12.—Monthly mean dissolved-solids loads at station 09217000 Green River near Green River, Wyoming.

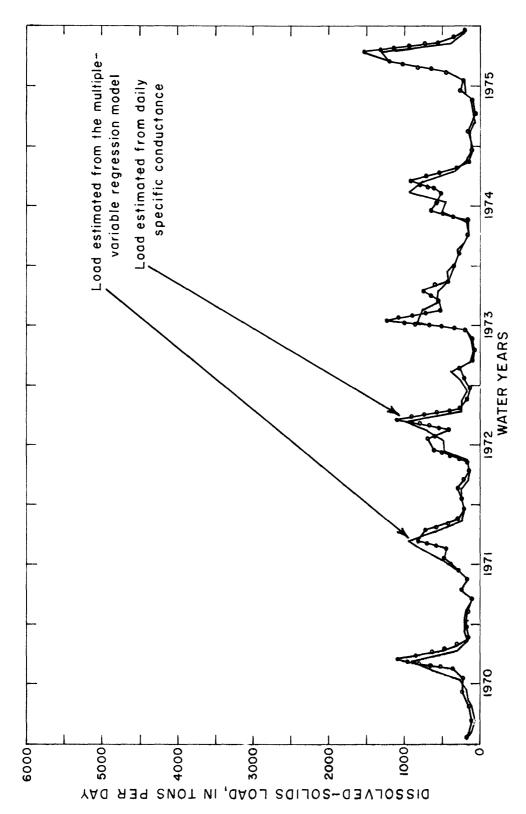


Figure 13.—Monthly mean dissolved-solids loads at station 09222000 Blacks Fork near Lyman, Wyoming.

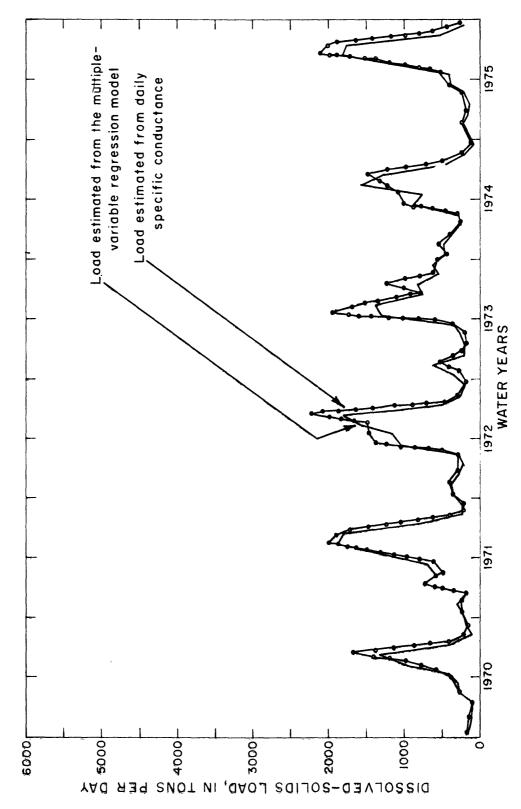


Figure 14.—Monthly mean dissolved-solids loads at station 09224700 Blacks Fork near Little America, Wyoming.

load hydrographs, comparison between stations of load magnitude and distribution with time can be made by visual inspection. For example, the base-flow loads at station 09211200 (fig. 10) increase in comparison to station 09209400 (fig. 9) without a corresponding increase in peak-flow loads. In contrast, comparisons between stations 09217000 (fig. 12) and 09211200 (fig. 10) indicate an increase of about 500 tons per day for both base-flow loads and peak-flow loads. While visual inspection of the example hydrographs aids in the evaluation of solute flow through a stream system, a more quantitative approach, as demonstrated in the following section, often is desirable.

Delineation of Sources of Salinity

Loads estimated at several points in a stream system collectively can provide quantitative information about the amount and chemical composition of dissolved solids gained in the intervening reaches.

For example, simulated dissolved-solids loads at stations 09211200 Green River below Fontenelle Reservoir and 09217000 Green River near Green River, Wyoming, show an average gain over the 1970-75 water years in the intervening reach of about 202,000 tons of dissolved solids per This gain represents about 33 percent of the load at station 09217000 and less than 5 percent of the streamflow. Big Sandy River is the major tributary to the Green River between stations 09211200 and 09217000. Simulated dissolved-solids loads averaged over the 1970-75 water years at station 09216000 Big Sandy River below Eden, Wyoming, 30 river miles upstream from the mouth, is 88,200 tons per year. The remaining increase of 114,000 tons per year is gained along the lower 30-mile reach of the Big Sandy River to the mouth and along the Green River between Fontenelle Reservoir and Green River, Wyoming. Cumulative dissolved-solids loads at stations 09211200, 09216000, and 09217000 are shown in figure 15 to illustrate the relative contribution of dissolved solids in the reaches between the stations.

More can be learned about the mean annual 114,000 ton-per-year dissolved-solids load by considering individual components of the load. Dissolved-sodium and dissolved-sulfate loads, plotted in figures 16 and 17, more than double in the Green River from below Fontenelle Reservoir to Green River, Wyoming. The average chemical composition of the 114,000 ton-per-year load is 84 percent sodium plus sulfate by weight compared to 31 and 72 percent sodium plus sulfate by weight in the loads at stations 09211200 and 09216000.

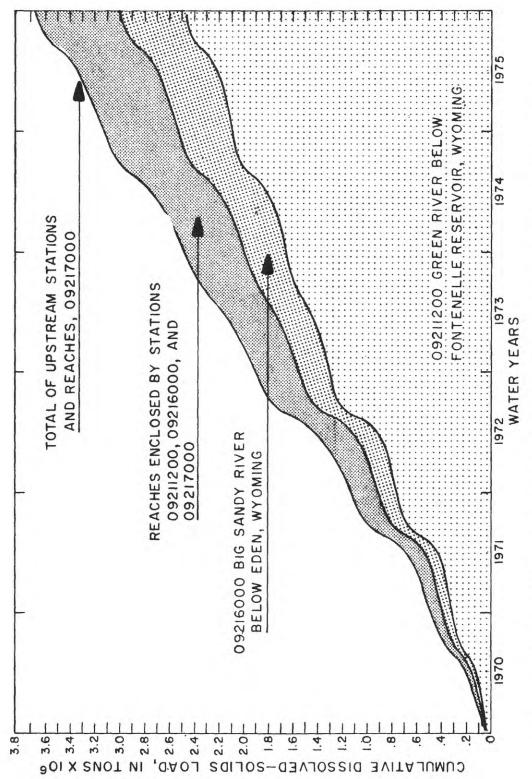


Figure 15.—Cumulative dissolved-solids loading in reaches enclosed by stations 09211200 Green River below Fontenelle Reservoir, Wyoming; 09216000 Big Sandy River below Eden, Wyoming; and 09217000 Green River near Green River, Wyoming.

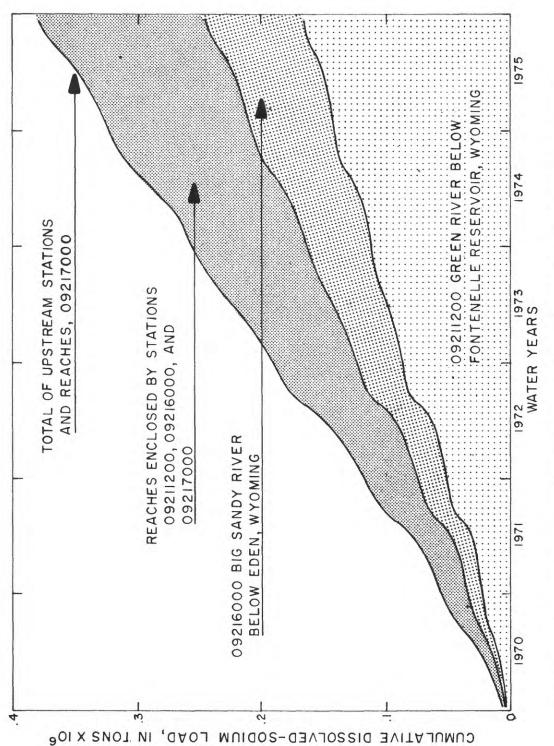


Figure 16.—Cumulative dissolved—sodium loading in reaches enclosed by stations 09211200 Green River below Fontenelle Reservoir, Wyoming; 09216000 Big Sandy River below Eden, Wyoming; and 09217000 Green River near Green River, Wyoming.

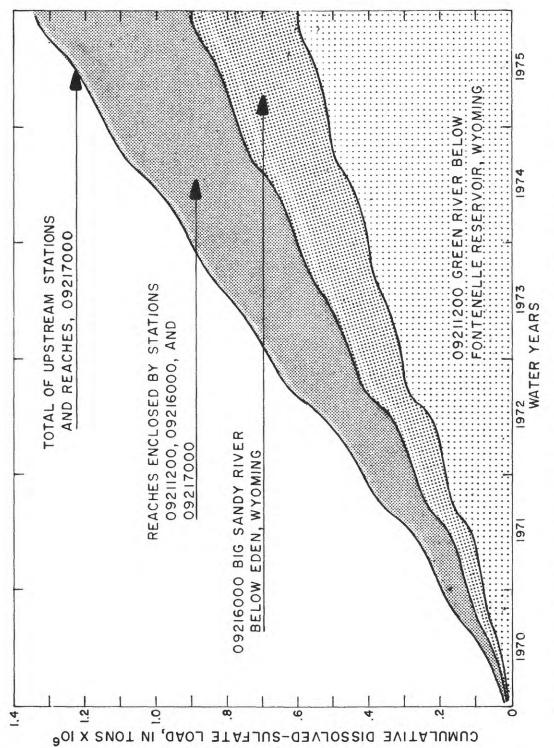
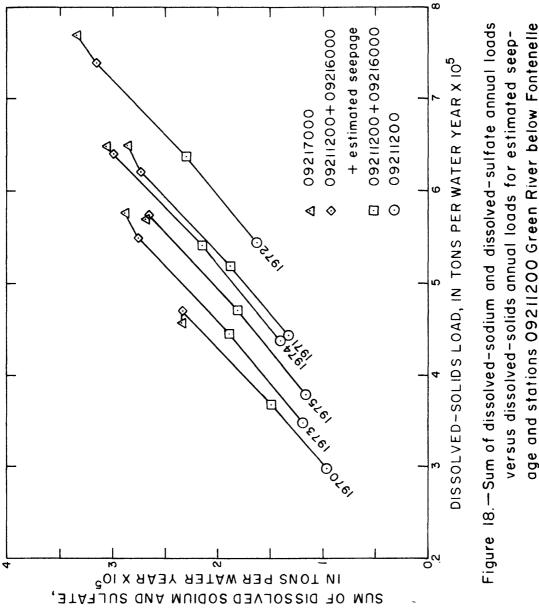


Figure 17.—Cumulative dissolved-sulfate loading in reaches enclosed by stations 09211200 Green River below Fontenelle Reservoir, Wyoming; 09216000 Big Sandy River below Eden, Wyoming; and 09217000 Green River near Green River, Wyoming.

The chemical character of the dissolved-solids gain serves as an indicator of sources. Samples from seeps along the Big Sandy River downstream from station 09216000 range from 3,800 to 6,800 milligrams per liter dissolved solids of which 84 percent is sodium plus sulfate by weight. Station 09216050 Big Sandy River at Gasson Bridge near Eden, Wyoming, (fig. 2) was established downstream from the seeps in May 1972 for the purpose of collecting streamflow records. Because water-quality sampling at the station was not initiated until February 1975, there are not yet enough data available to estimate dissolved-solids loads to quantitatively determine the dissolved-solids contribution of seeps along the Big Sandy River between stations 09216000 and 09216050. Discharge from the seeps has not been measured directly, but streamflow records at the two stations indicate a mean discharge gain of about 20 cubic feet per second in October when there is less evapotranspiration and negligible surface-water gain. Based on an average flow from the seeps of 20 cubic feet per second at a concentration of 5,000 milligrams per liter dissolved solids, the annual discharge from the seeps would average about 100,000 tons of dissolved solids which would account for about 88 percent of the load gained in the Green River and Big Sandy reaches enclosed by stations 09211200 and 09216000 upstream, and 09217000 downstream. To demonstrate how the amount and chemical character collectively aid in delineating sources of salinity, the sum of sodium and sulfate loads versus total dissolved-solids load is plotted for this example in figure 18. The close proximity of points representing the sum of the estimated seepage load and upstream stations to points representing loads at station 09217000 indicates good agreement both in amount and chemical character of the load gained in the intervening reaches despite a relatively large variation in loads at stations 09211200 and 09216000. Analyses similar to this example can be used in many other reaches where discrete monthly samples and daily streamflow records are available. This type of analysis would be difficult based on discrete monthly samples alone.



age and stations 09211200 Green River below Fontenelle Eden, Wyoming; and 09217000 Green River near Green Reservoir, Wyoming; 09216000 Big Sandy River below River, Wyoming.

SUMMARY

Daily concentration of dissolved solids in a stream may be estimated from daily streamflow records using a multiple-variable regression model developed from chemical analyses of samples collected on a monthly basis. The model relates dissolved-solids concentration of the stream to stream-Seasonal variation of dissolved solids not directly related to streamflow are accounted for in the model by the incorporation of harmonic functions of time. Because of the variability in streamflow and dissolvedsolids concentration of streams in the Green River Basin, monthly mean loads and concentrations computed from daily estimates from the model provide a better representation of overall dissolved-solids concentration of the streams than do discrete monthly samples. Consequently, estimates from the model of dissolved-solids concentrations provide information useful to water planners and managers concerned with the evaluation of impacts of proposed and past water-development projects (such as reservoirs irrigation systems, and withdrawals for municipal and industrial use). The model may also be utilized in assessing the feasibility of reduced sampling frequencies for providing continuing information on long-term trends in salinity-streamflow relations and shifts in sampling locations for providing additional information on sources of salinity. An overall reduction in the data collection effort allocated to salinity in the streams of the Green River Basin would allow greater emphasis to be applied to other equally important water-quality factors for which few data are presently available.

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Table 2.--Sequence number conversion, date to water-year day

Day	0ct	Nov	Dec	Jan	Feb	Mar	Apr	May	June	July	Aug	Sept
1	1	32	62	93	124	152	183	213	244	274	305	336
2	2	33	63	94	125	153	184	214	245	275	306	337
3	3	34	64	95	126	154	185	215	246	276	307	338
4	4	35	65	96	127	155	186	216	247	277	308	339
5	5	36	66	97	128	156	187	217	248	278	309	340
,	,	50	00	,,	140	130	407		2-10	270	303	340
6	6	37	67	98	128	157	188	218	249	279	310	341
7	7	38	68	99	130	158	189	219	250	280	311	342
8	8	39	69	100	131	159	190	220	251	281	312	343
9	9	40	70	101	132	160	191	221	252	282	313	344
10	10	41	71	102	133	161	192	222	253	283	314	345
		,				04			<u>-</u>			
11	11	42	72	103	134	162	193	223	254	284	315	346
12	12	43	73	104	135	163	194	224	25 5	285	316	347
13	13	44	74	105	136	164	195	225	256	286	317	348
14	14	45	75	106	137	165	196	226	257	287	318	349
15	15	46	76	107	138	166	197	227	258	288	319	350
							-					-
16	16	47	77	108	139	167	198	228	259	289	320	351
17	17	48	78	109	140	168	199	229	260	290	321	352
18	18	49	79	110	141	169	200	230	261	291	322	353
19	19	50	80	111	142	170	201	231	262	292	3 23	354
20	20	51	81	112	143	171	202	232	263	293	324	355
21	21	52	82	113	144	172	203	233	264	294	325	356
22	22	53	83	114	145	173	204	234	265	295	326	35 7
23	23	54	84	115	146	174	205	2 35	266	296	327	358
24	24	5 5	85	116	147	175	206	236	267	297	328	359
25	25	56	86	117	148	176	207	237	268	298	329	360
						_						
26	26	57	87	118	149	177	208	238	269	299	330	361
27	27	58	88	119	150	178	209	239	270	300	331	362
28	28	59	89	120	151	179	210	240	271	301	332	363
29	29	60	90	121	(152)	180	211	241	272	302	333	364
30	30	61	91	122		181	212	242	273	303	334	365
31	31		92	123		182		243		304	335	

Note: For months of March through September add one (1) to number in table for sequence conversion of days for leap years.

Table 3.--Regression results, concentration versus discharge and time

 $[B_0 + B_1 \sin(\alpha t) + B_2 \cos(\alpha t)] [B_3 + B_4 \sin(\alpha t) + B_5 \cos(\alpha t)]$ C = 10

where

C = Constituent concentration, in milligrams per Liter. Q = Discharge, in cubic feet per second. Q = Discharge, Q = Discharge, in cubic feet per second. Q = Discharge, Q = Discharge

t = Day of water year.

Constituents (concentrations are in milligrams per liter):

Ca = calcium $\frac{1}{2}$ HCO3 = bicarbonate $\frac{1}{2}$ Mg = magnesium $\frac{1}{2}$ SO4 = sulfate $\frac{1}{2}$ CL = chloride $\frac{1}{2}$ K = potasium $\frac{1}{2}$ TDS = dissolved solids

Con- stitu- ent	Во	В1	В2	В 3	Вц	B ₅	r	SE (log units)	N
	09188	500 Green Ri	ver at Warren	Bridge, near Da	aniel, Wyo., 1	968-75 water ye	ears		
Ca	2.4454	-0.3368	0.5140	-0.2555	0.1819	-0.2513	0.981	0.045	78
Mg	1.8932	3916	.6109	3043	.2072	3014	.949	.086	78
Na	.8680	4383	.5463	1274	.2191	2654	.829	.116	78
K	.0586	4819	.2751	.9698X10 ⁻²	.2265	1690	.695	.128	78
HCO ₃	2.4140	2260	.6311	1248	.1293	3068	.965	.044	78
SO4	3.0835	5370	.3708	4159	.2739	1863	.986	.055	77
C1	.3870	3415	.6011	8663X10 ⁻¹	.1641	2640	.415	.250	77
TDS	3.0855	3275	.4789	2617	.1814	2407	.985	.041	78
	091926	600 Green Ri	ver near Big P	iney, Wyo., 196	67-75 water yea	ars			
Ca	1.7843	0.9890X10 ⁻¹	0.9982X10 ⁻¹	-0.3792X10 ⁻²	0.2161X10 ⁻¹	-0.5733X10 ⁻¹	0.910	0.056	89
Mg	1.0344	$7263X10^{-1}$.5276X10 ⁻¹	.5451x10 ⁻¹	.9910X10 ⁻¹	5017X10 ⁻¹	.825	.095	89
Na	.4995	9178	.3005	.1697	.3824	1512	.569	.164	89
K	.2784	3790	.3183	.1800	.1863	1610	.712	.102	89
HCO ₃	1.9968	.9484X10 ⁻¹	.1246	.8842X10 ⁻¹	.1397X10 ⁻¹	7213X10 ⁻¹	.852	.053	89
SO _L	2.2344	4376	.4258X10-1	1120	.2469	4056X10 ⁻¹	.945	.080	89
C1	.9049	2238	.6090	1906	.7529X10- ¹	÷ .2718	.314	.339	85
TDS	2.4224	7516X10 ⁻¹	.1077	.5206x10 ⁻²	.8699X10 ⁻¹	6815X10 ⁻¹	.927	.052	89
	092050	000 New For	k River near B	ig Piney, Wyo.,	, 1969-75 water	year			
Ca	2.1925	-0.4160	-0.4160	-0.2947	0.1498	0.1200	0.893	0.078	71
Mg	.7144	5475	5633	1464X10 ⁻¹	.2410	.2187	.564	.186	70
Na	1,4866	8208	2626	1899	.3111	.1078	.858	.098	71
K		6029	8019X10 ⁻¹	4406X10 ⁻¹	.2294	.2738X10 ⁻¹	.567	.120	70
HCO ₃	2.7188	4023	5371	2641	.1458	.2191	.879	.083	71
S04	1.5564	7377	.3968	1672	.2992	1518	.631	.196	70
C1	.7356	7806	.3370	8435X10 ⁻¹	.3120	1411	.532	.200	68
TDS	2.6440	4413	2889	- ,2109	.1687	.1254	.851	.087	70
	0920 9 4	00 Green Riv	ver near La Bar	ge, Wyo., 1970	-75 water year	s			
Са	1.9262	-0.1408	0.4260	-0.8631X10 ⁻¹	0.6860X10 ⁻¹	-0.1532	0.857	0.051	59
Mg		1518	.1662	1294	.6249X10 ⁻¹	7216X10 ⁻¹	.644	.100	59
Na	1.3806	-1.150	.6504	6527X10 ⁻¹	.4034	2239	.786	.107	59
K		6747	.7754	.4045X10 ⁻¹	.2318	2802	.613	.104	59
HCO		1116	.3261	9231X10 ⁻¹	.4692X10 ⁻¹	1192	.784	.049	58
SO		7708	.4432	1699	.2946	- ,1665	.926	.075	58
01		-1.028	,3341	.1219	.3955	1445	.620	.187	59
TDS		4033	.3956	6532X10 ⁻¹	.1555	1459	.864	.053	59
	092112	00 Green Riv	ver below Fonte	nelle Reservoi	r, Wyo., 1970-	75 water years			
Ca	1.8813	-0.7463X10 ⁻¹	0.1242	-0.6576X10 ⁻¹	0.3601X10 ⁻¹	-0.6532X10 ⁻¹	0.858	0.044	59
Mg	2.1434	.3819	9552X10 ⁻¹	3400	1351	$8903x10^{-3}$.582	.118	59
Na		7561X10 ⁻¹	.5277	2969	.3505x10 ⁻¹	1969	.838	.079	59
K	,2186	,1826	2944	7194X10 ⁻²	$6705X10^{-2}$.7906X10 ⁻¹	.494	.081	59
HCO ₃	2,3076	.1433X10 ⁻²	7979X10 ⁻¹	3280x10 ⁻¹	.4625X10-2	.5705x10 ⁻²	.785	.039	59
SO ₄		2999	.8828	3392	.1028	3247	.897	,074	56
C1	1.1211	.9124X10 ⁻¹	4148	1548	9964X10 ⁻²	.7716X10 ⁻¹	,660	.172	58
TDS	2,9542	,4108X10 ⁻¹	.2386	1848	8943X10 ⁻²	1036	.904	.038	59
	£+7J4£	*4100VIO	.2300	.1040	.0774410	+1000	• > 0 -	.030	

Table 3.--Regression results--continued

Con- stitu- ent	Во	В	B ₂	Вз	Вц	B ₅	r	SE (log units)	N
	09216	0000 Big Sandy	River below I	Eden, Wyo., 19	61-75 water yea	ars			
Ca	2.8308	-0.2819	-0.3118	-0.3193	0.2238	0.2298	0.930	0.065	135
Mg	2.5163	4769	4435	3962	.3409	.2906	.763	.170	133
Na	3.1705	5360	1425	4107	.3546	.9222X10 ⁻¹	.921	.075	136
K	.8432	2324	1871	1831	.1371	.1271	.747	.070	134
HCO ₃	2.6226	2080	2263	1526	.1509	.1526	.856	.052	136
SO4	3.7132	4768	3024	3867	.3453	.2078	.939	.071	132
C1	2.6059	6753	5450×10^{-1}	5222	.3899	.5756x10 ⁻¹	.911	.090	134
TDS	3.9191	4373	2577	3697 	.3102	.1791	.938	.065	134
	09216	300 Green Riv	er at Big Isla	and, near Gree	n River, Wyo.,	1966-75 water	years		
Ca	2.3985	0.9136x10 ⁻¹	0.1184	-0.2114	-0.1983X10 ⁻¹	-0.5649X10 ⁻¹	0.906	0.042	102
Mg	2.3717	.1850	.118	3557	5387X10 ⁻¹	5183X10 ⁻¹	.789	.097	102
Na	3.5144	1151	.4623	6195	.3199X10 ⁻¹		.900	.094	102
K	.4599	.4617X10 ⁻¹	.1556	7425X10 ⁻¹	2488X10 ⁻¹		.298	.092	102
HCO ₃	2.4754	.2112	$8044X10^{-2}$	8492X10 ⁻¹	5954X10 ⁻¹	1613×10^{-1}	.843	.038	102
S04	4.1100	2129	.3763	6234	.7070X10 ⁻¹	1364	.921	.088	101
C1	2.7664	6385	.6583	6096	.1907	2426	.804	.141	101
TDS	2.4358	.1564X10 ⁻¹	.5281X10 ⁻¹	6429X10 ⁻¹	.8525X10 ⁻³	2016X10 ⁻¹	.981	.026	90
	09217	000 Green Rive	r near Green R	liver, Wyo., 19	969-75 water ye	ears			
Ca	2.3422	.7921X10 ⁻¹	.3339	3339	1773X10 ⁻¹	1223	.914	.031	79
Mg	2.0169	$6733X10^{-1}$.5429X10 ⁻¹	2299	.3125X10 ⁻¹	3226X10 ⁻¹	.849	.053	79
Na	2.9491	7523	.1445	4147	.2477	5830X10 ⁻¹	.833	.105	79
K	.1949	3300	.1001	.2950X10 ⁻¹	.1050	4980X10 ⁻¹	.373	.114	68
HCO ₃	2.4509	.1540	.1388	7179×10^{-1}	4039X10 ⁻¹	6137X10 ⁻¹	.866	.029	79
S0 ₄	3.7297	4682	.4333	4920	.1560	1485	.934	.065	68
C1	1.8550	4853	.3897	3137	.1587	1463	.802	.089	68
TDS	3.4821	1001	.3078	2910	.4131X10 ⁻¹	1087	.935	.039	6 8
	09222	000 Blacks Fo	rk near Lyman,	Wyo., 1970-7	5 water years				
Ca	2.6737	-0.1666	0.2548X10 ⁻¹	-0.3078	0.2669X10 ⁻¹	0.3359x10 ⁻¹	0.951	0.062	53
Mg	2.2836	3277	3364X10 ⁻¹	3487	.1007	.7367x10 ⁻¹	.934	.089	54
Na	2.6877	6437	.3100	2599	.2574	9781x10 ⁻¹	.884	.126	53
K	.5752	4374	.2794	4799X10 ⁻¹		1409	.775	.082	54
HCO ₃	2.5942	9595X10 ⁻³	2064	1076	1372X10 ⁻¹	.9670x10 ⁻¹	.679	.053	54
SO4	3.3881	4991	.1929	3596	.1776	1158x10 ⁻¹	.931	.113	54
C1	1.9935	7167	.2139	1367	.3011	8693X10 ⁻¹	.710	.167	54
TDS	3.5611	4165	.1823	2872	.1476	3277X10 ⁻¹	.930	.087	54
	09224	450 Hams Fork	near Granger,	Wyo., 1969-7	5 water vears				
						0. 5120210=1	0.771	0.000	
Са	1.9122 1.6777	0.1702 .1173	-0.1251	-0.3613x10 ⁻¹	-0.6055X10 ⁻¹	0.5120X10 ⁻¹	0.771	0.069	88
Mg Na			1355	1912	7256X10 ⁻¹	.6406x10 ⁻¹	.815	0.084	88
na K	1.9557 .2819	.1579X10-1 .9825X10-1	1859	2271	5991X10 ⁻²	.1140	.832	.116	88
HCO ₃	2.4167	.1674	.1170X10 ⁻¹ 9874X10 ⁻¹	2305X10 ⁻¹	5176X10 ⁻¹	3552X10 ⁻¹	.476	.082	88
нсо _з SO ₄	2.4167	.16/4 .2888X10 ⁻¹		5596X10 ⁻¹	1046	.4351x10-1	.658	.057	88
50 ₄ C1	1.4965	.5568X10 ⁻¹	2209 2590	2006 1467	.3031X10 ⁻¹	.1248	.890	.103	86
TDS	2.8561	.1368	2390 1227	1392	5930X10 ⁻¹	.1355 .5637x10 ⁻¹	.845 .855	.127 .066	88 88
					., 1970-75 wate		.033		
Ca	2.1897	-0.3005X10 ⁻¹	-0.2452	-0.8708X10 ⁻¹	0.1222X10 ⁻¹	0.1432	0.778	0.087	61
Mg	1.9720	3323	3435	1506	.1459	.1844	.894	.082	62
Na 	2.5947	6160	4195	1768	.2907	.2250	.817	.168	62
K	.7172	3418	3822	3461X10 ⁻¹	.1374	.1560	.661	.078	55
HCO ₃	2.5747	.2151	3399	7689X10 ⁻¹	8641X10 ⁻¹	.1328	.759	.061	62
SO ₄	3.2412	4122	2117	2758	.1868	.1505	.903	.115	54
C1	2.1541	5409	3045	1707	.2657	.1348	.770	.153	55
TDS	3.4221	2466	1394	2064	.1164	.9722X10 ⁻¹	.886	.092	55
	09259	700 Little Sn	ake River near	Baggs, Wyo.,	1965-74 water	years			
Ca	1.6052	-0.3896	-0.4413X10 ⁻¹	0.8984X10 ⁻²	0.2174	0.2892X10 ⁻¹	0.895	0.078	81
Мg	1.1950	7907	1716	4597X10 ⁻¹	.3869	.7229x10-1	.844	.177	81
Na	1.9928	8096	.1733	2087	.3709	5973X10 ⁻¹	.952	.134	81
K	.4759	4543	.6136x10 ⁻¹	5436X10 ⁻¹	.2002	3844X10 ⁻¹	.792	.126	80
HCO ₃	2.4537	2912	1615	8414X10 ⁻¹	.1598	.8360X10 ⁻¹	.952	.061	81
SO ₄	1.9059	-1.027	.3184	4803X10 ⁻¹	.4940	1370	.920	.156	80
C1	1.7546	6815	.2016	3413	.2858	5831X10 ⁻¹	.924	.203	79
TDS	2.704	5203	.1653	1190	.2498	7013X10 ⁻¹	.961	.075	80

Table 4.--Regression results, concentration versus specific conductance

[S. J. Rucker IV, written commun., 1977]

TDS = E+FK

where

TDS = Dissolved solids, in milligrams per liter,

E = Intercept, in milligrams per liter,

F = Slope, and

K = Specific conductance, in micromhos per centimeter
at 25°C.

Station	E	F	r	SE	N
09209400	-12.9	0.645	0.987	9.6	129
09211200	-23.6	.657	.942	10.4	83
09217000	-57.5	.760	.993	21.1	149
09222000	-18.4	.856	.993	86.3	150
09224700	-88.8	.772	.995	52.6	154

r = Correlation coefficient.

SE = Standard error of estimate, in milligrams per liter.

N = Number of paired values.

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